

# Prenatal Exposure to Low Doses of Bisphenol A Increases Pituitary Proliferation and Gonadotroph Number in Female Mice Offspring at Birth<sup>1</sup>

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## ABSTRACT

The pituitary gland is composed of hormone-producing cells essential for homeostasis and reproduction. Pituitary cells are sensitive to endocrine feedback in the adult and can have altered hormonal secretion from exposure to the endocrine disruptor bisphenol A (BPA). BPA is a prevalent plasticizer used in food and beverage containers, leading to widespread human exposure. Although prenatal exposure to BPA can impact reproductive function in the adult, the effects of BPA on the developing pituitary are unknown. We hypothesized that prenatal exposure to low doses of BPA impacts gonadotroph cell number or parameters of hormone synthesis. To test this, pregnant mice were administered 0.5 µg/kg/day of BPA, 50 µg/kg/day of BPA, or vehicle beginning on Embryonic Day 10.5. At parturition, pituitaries from female offspring exposed in utero to either dose of BPA had increased proliferation, as assessed by *mKi67* mRNA levels and immunohistochemistry. Coincidentally, gonadotroph number also increased in treated females. However, we observed a dichotomy between mRNA levels of *Lhb* and *Fshb*. Female mice exposed to 0.5 µg/kg/day BPA had increased mRNA levels of gonadotropins and the gonadotropin-receptor hormone (GNRH) receptor (*Gnrhr*), which mediates GNRH regulation of gonadotropin production and release. In contrast, mice treated with 50 µg/kg/day of BPA had decreased gonadotropin mRNA levels, *Gnrhr* and *Nr5a1*, a transcription factor required for gonadotroph differentiation. No other pituitary hormones were altered on the day of birth in response to in utero BPA exposure, and male pituitaries showed no change in the parameters tested. Collectively, these results show that prenatal exposure to BPA affects pituitary gonadotroph development in females.

*anterior pituitary, bisphenol A (BPA), development, developmental biology, endocrine disruptors, environmental contaminants and toxicants, follicle-stimulating hormone (FSH), gonadotrophs, gonadotropins, luteinizing hormone (LH), pituitary*

<sup>1</sup>Supported by National Institutes of Health grant R01DK076647 (L.T.R.), P20ES018163 (J.A.F), University of Illinois at Urbana-Champaign Research Board Grant 12174 (L.T.R.), and the Billie A. Field Fellowship (W.W. and Z.R.C).

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Received: 15 March 2012.

First decision: 9 May 2012.

Accepted: 1 August 2012.

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eISSN: 1529-7268 <http://www.biolreprod.org>

ISSN: 0006-3363

## INTRODUCTION

The pituitary gland develops from an invagination of the oral ectoderm during embryonic development to form the structure known as the Rathke pouch. This structure contains highly proliferative progenitor cells that receive signals from the neighboring neural and oral ectoderm to initiate transcription factor expression in a spatial and temporal manner [1–3]. This process is necessary for differentiation of the distinct hormone-producing cell types [4]. The extent to which cell specification is hardwired, or the ability of these signals to be modified by the in utero environment, is currently unknown.

Gestational exposure to chemicals that mimic or inhibit hormone action, commonly known as endocrine-disrupting chemicals (EDCs), can interfere with the hypothalamic-pituitary-gonadal (HPG) axis and lead to infertility or disease in adults [5]. One of the most abundant EDCs is bisphenol A (BPA), which has been found in 95% of adult human urine samples tested and can accumulate in reproductive organs [6]. The acceptable human intake of BPA is referred to as the oral reference dose (ORfD) and is calculated by the U.S. Environmental Protection Agency to be 50 µg/kg/day [7]. Additionally, BPA has the ability to cross the placental barrier, but the fetus has limited abilities to metabolize it, leading to a high accumulation of BPA in fetal and placental tissue to levels that are equal to or greater than the ORfD [8].

Exposure to BPA has adverse effects on reproductive processes in both male and female rodents. Both embryonic and neonatal exposure to BPA at low doses can result in permanent reductions in fertility and fecundity, accompanied by advanced pubertal onset and abnormal estrous cyclicity [9–11]. Disruption has been documented at all levels of the HPG axis. The sexually dimorphic differences seen in the hypothalamic anteroventral periventricular nucleus (AVPV), critical for the luteinizing hormone (LH) surge, are diminished by BPA exposure [12]. Additionally, low dose in utero exposure can alter the morphology of the uterus, decrease the number of ovarian follicles in females, and decrease testis size and reduce sperm count in males [13–15]. At the level of the pituitary, embryonic or postnatal administration of BPA leads to decreased gonadotropin-receptor hormone (GNRH)-induced LH secretion in adults [10, 16]. Similarly, prepubertal BPA exposure also causes a reduction of serum LH and *Lhb* mRNA in males and females [17–19]. Although these data suggest that BPA exposure might impact pituitary physiology, nothing is known about the effect of BPA exposure on development of gonadotrophs, the cell type that synthesizes and secretes LH and follicle-stimulating hormone (FSH).

The process of gonadotroph specification requires a sequential activation of transcription factors within the developing pituitary. A key initiating event is the expression of *Pitx2* and *Gata2* and the absence of *Pit1* [20, 21]. For example, in mice with pituitary-specific *Gata2* loss of function,

TABLE 1. List of primer sequences.

Gene	Forward primer	Reverse primer
<i>Gapdh</i>	5'-GGT GAC GCC GGT GCT GAG TAT G-3'	5'-GAC CCG TTT GGC TCC ACC CTT C-3'
<i>Lhb</i>	5'-CCC AGT CTG CAT CAC CTT CAC-3'	5'-GAG GCA CAG GAG GCA AAG C-3'
<i>Fshb</i>	5'-TGG TGT GCG GGC TAC TGC TAC-3'	5'-ACA GCC AGG CAA TCT TAC GGT CTC-3'
<i>Ch</i>	5'-AGG GCA TCC AGG CTC TGA T-3'	5'-GCA TGT TGG CGT CAA ACT TG-3'
<i>Tshb</i>	5'-GCC GTC CTC CTC TCC GTG CTT-3'	5'-AGT TGG TTC TGA CAG CCT CGT G-3'
<i>Pomc</i>	5'-GTT ACG GTG GCT TCA TGA CCT C-3'	5'-CGC GTT CTT GAT GAT GGC GTT C-3'
<i>Prl</i>	5'-TCA GCC CAG AAA GCA GGG ACA-3'	5'-GGC AGT CAC CAG CGG AAC AGA-3'
<i>Nr5a1</i>	5'-AAA TTC CTG AAC AAC CAC AGC-3'	5'-GCA TCT CAA TGA GAA GGT TG-3'
<i>Gnrhr</i>	5'-ATG ATG GTG GTG ATT AGC C-3'	5'-ATT GCG AGA AGA CTG TGG-3'
<i>Egr1</i>	5'-GGA GTG ATG AAC GCA AGA G-3'	5'-AGC CAG GAG GGA GTA GC-3'
<i>Mki67</i>	5'-CCA GGG ATC TCA GCG CAA TTA CAG-3'	5'-GGA TAG GAC AGA GGG CCA CAT TTC-3'

gonadotroph number is severely decreased, and the circulating levels of LH and FSH are reduced [22]. Additionally, mice with decreased pituitary expression of *Pitx2* have reduced expansion of the Rathke pouch early in development and fail to activate several transcription factors critical for gonadotroph differentiation. This results in diminished or complete absence of *Lhb* and *Fshb* depending on the severity of the mutation [20, 23]. Subsequently, the transcription factors *Egr1* and *Nr5a1* (*Sfl*) are key components of initiating either *Lhb* or both *Lhb* and *Fshb* transcription, respectively, during development [24–26]. It is currently unknown if in utero exposure to EDCs can impact these initial stages of gonadotroph formation or gonadotroph cell number.

Based on the ability of embryonic exposure to BPA to disrupt HPG axis function and the ability of BPA to accumulate in fetal tissue [8], we hypothesized that prenatal exposure to BPA at or below the ORfD results in changes in pituitary development, especially in gonadotrophs. The present study was designed to examine the impact of maternal BPA exposure on proliferation and cell specification in the pituitary of male and female offspring at birth.

## MATERIALS AND METHODS

### Animals

Timed pregnancies of mice on a mixed FVB, C57BL/6 background were generated. The morning after a vaginal plug was observed was designated as e0.5 and the day of parturition was designated as P1. Pregnant females were dosed orally once a day with either 0.5 or 50  $\mu\text{g}/\text{kg}/\text{day}$  of 2,2-bis(4-hydroxyphenyl)propane-4,4'-isopropylidenediphenol (BPA) dissolved in ethanol and diluted in tocopherol-stripped corn oil or with tocopherol-stripped corn oil alone as a control. Mice were dosed from e10.5 through e18.5. Mice born to treated mothers were euthanized on P1 for experiments. For each treatment group, from six to eight individual pituitaries were examined. Pups were collected from five to seven different litters per treatment group. Mice used for characterization of the onset of prolactin (PRL) expression were from a CD-1 background taken at P1, P5, P10, and 4–6 mo of age, with six biological replicates for each age. All animals were housed in a facility with a 12L:12D photoperiod and fed a standard mouse diet. The University of Illinois at Urbana-Champaign Institutional Animal Care and Use Committee approved all procedures.

### Immunohistochemistry

Heads from P1 mice were fixed overnight in 3.7% formaldehyde in PBS (pH 7.2) at 4°C, dehydrated in a graded series of ethanol at room temperature, and embedded in paraffin for immunohistochemistry. Paraffin blocks were sectioned (thickness, 6  $\mu\text{m}$ ) and prepared for immunohistochemistry as previously described [27].

Pituitary cell populations were examined with antibodies against LH $\beta$  (1:1000, AFP22238790GPOLHB; National Hormone and Pituitary Program [NHPP], National Institute of Diabetes and Digestive and Kidney Diseases), FSH $\beta$  (1:1000, 85GP9691bFSHB; NHPP), PRL (1:1000, lot AFP425\_10\_91; NHPP), and mKi67 (1:100; Dako). Biotin-conjugated secondary antibodies were used (1:250) and amplified with Streptavidin-Cy3 (1:250; Jackson

ImmunoResearch Laboratories, Inc.). All experiments contained a slide processed without primary antibody as a control.

For colocalization experiments, blocking and incubation of the primary antibody occurred as above, with mKi67 (1:100; BD Pharmingen), FSH $\beta$  (1:1000, 85GP9691bFSHB; NHPP), SOX2 (1:250; Millipore), and PIT1 (1:500; a gift from Dr. Simon Rhodes, Indiana University, Purdue University, Indianapolis). An anti-mouse secondary antibody conjugated to biotin was used with mKi67, whereas anti-rabbit secondary antibody conjugated to Cy3 fluorophore was used with FSH $\beta$ , SOX2, and PIT1. Streptavidin-conjugated DyLight 488 fluorophore was used as a tertiary antibody to detect the anti-mouse biotin. All secondary and streptavidin-conjugated antibodies were purchased from Jackson ImmunoResearch and were used at a concentration of 1:200. Controls for these experiments included a slide processed without mKi67 and a slide processed without FSH, PIT1, or SOX2 antibodies.

All slides were counterstained with 4',6-diamidino-2-phenylindole (DAPI; 1:1000; Sigma) and visualized at 100 $\times$  and 400 $\times$  magnification using a Leica DM2560 microscope. Photographs were taken using a Retiga 2000R camera (Q-Imaging) and acquired using Q-Capture Pro software (Q-Imaging). Images were processed using Adobe Photoshop CS2.

### Quantitative RT-PCR

Pituitaries isolated from individual P1 mice were stored in RNAlater (Ambion) at  $-20^\circ\text{C}$ . RNAlater was removed, and RNA was isolated from individual pituitaries using an RNAqueous-Micro Kit (Ambion) as per the manufacturer's protocol. All cDNA was synthesized from 6  $\mu\text{l}$  of RNA using the ProtoScript M-MuLV First Strand cDNA Synthesis Kit (New England Biolabs). A no-enzyme control was also prepared and used as a negative control. Pituitary cDNA was amplified using gene-specific primers (Table 1) and SYBR green mix (Bio-Rad Laboratories) on a Bio-Rad iQ5 real-time PCR machine. Data were analyzed with the standard comparative ( $\Delta\Delta\text{Ct}$ ) method. For each sample, the mean Ct for the gene of interest and for the control gene, *Gapdh*, was calculated as an average of the duplicates of that sample. The  $\Delta\text{Ct}$  was calculated by subtracting the mean *Gapdh* Ct value from the mean gene of interest Ct value. The  $\Delta\Delta\text{Ct}$  was calculated as the difference between the  $\Delta\text{Ct}$  between the treatment groups and the vehicle control groups, or the vehicle control groups and the vehicle control groups to give the normalized level of target gene. The relative fold-change of expression was then equaled to  $2^{(-\Delta\Delta\text{Ct})}$  for each sample. The error bars in the figures represent the SEM of the relative fold-change for each group. A sample size of from six to eight individual pituitaries, run in duplicate, was used for all groups. Pups were collected from five to seven different litters per treatment group.

### Cell Quantification

To quantify the number of gonadotrophs present in the pituitary, four animals per treatment group were used. For each animal, a total of five slides was chosen approximately 70  $\mu\text{m}$  apart and spanning the entirety of the pituitary gland (360  $\mu\text{m}$  in total). Each slide contained two sections (thickness, 6  $\mu\text{m}$ ). The number of LH $\beta$ - or FSH $\beta$ -immunopositive cells were counted and divided by the total number of DAPI-positive cells. Cell quantification data are shown as the average percentage of LH $\beta$ - or FSH $\beta$ -immunopositive cells over the average number of total cells throughout the pituitary for each group.

### Statistical Analysis

Statistical significance was determined using ANOVA and subsequent comparisons using a two-tailed *t*-test in StatPlus (AnalystSoft) and Microsoft

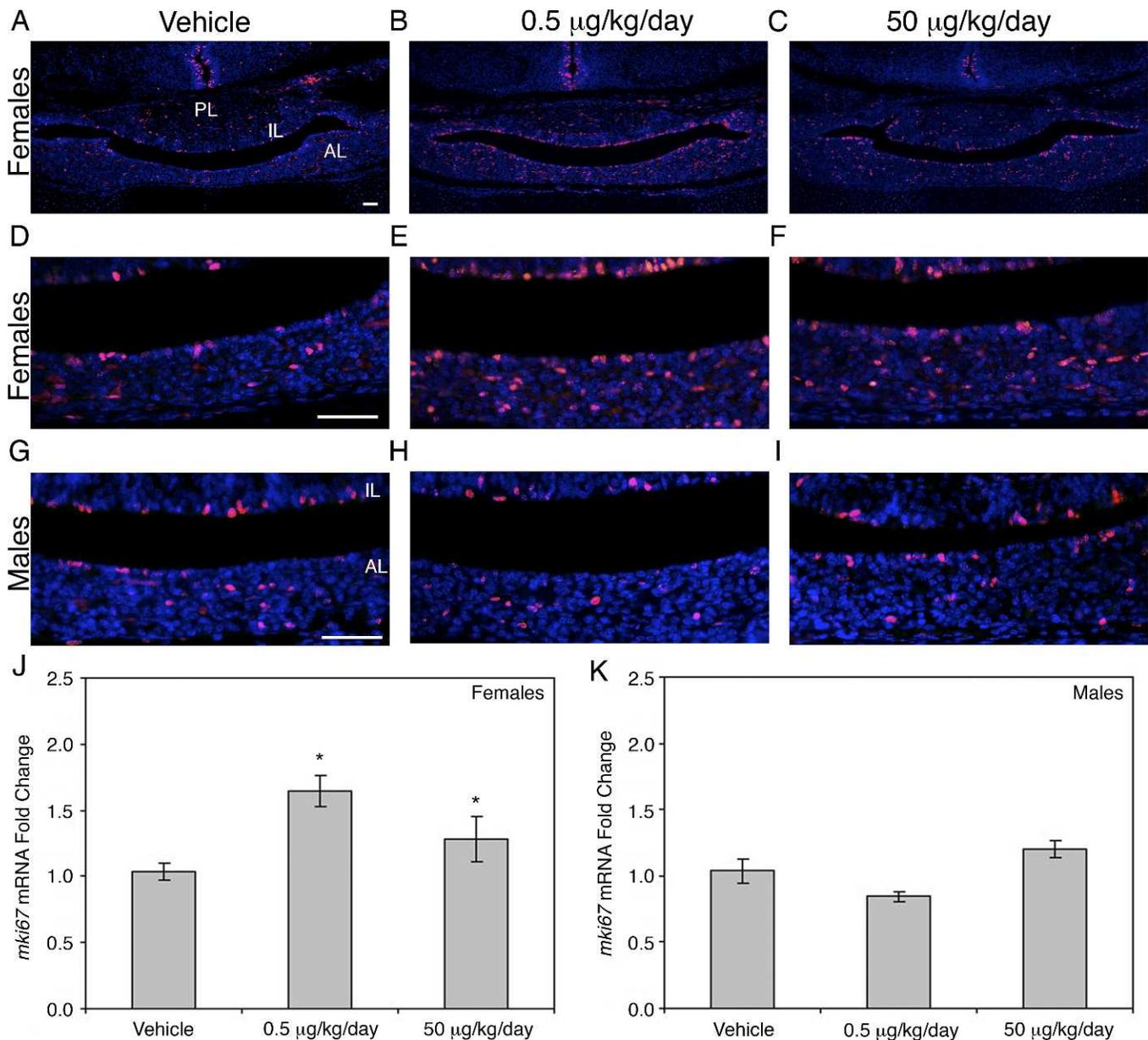


FIG. 1. BPA exposure increased proliferation in the pituitary of females. The number of mKi67-immunoreactive cells appeared to be greater in the female pituitaries of both the 0.5 µg/kg/day dose (B and E) and the 50 µg/kg/day dose (C and F) of BPA treatment groups compared to the vehicle control (A and D). In males, there appeared to be no difference in mKi67-immunoreactive cells for any condition (G–I). *mki67* mRNA levels were significantly increased in females of both treatment groups (J;  $P=0.0005$ , 0.5 µg/kg/day;  $P=0.0375$ , 50 µg/kg/day). *mki67* mRNA levels in males showed no change compared to the vehicle control group (K). All qRT-PCR values are normalized to the housekeeping gene *Gapdh*, and values are represented relative to the vehicle control group. Significance in J and K is noted by an asterisk ( $*P \leq 0.05$ ,  $n = 4$  [immunohistochemistry],  $n = 6-8$  [qRT-PCR]). AL, anterior lobe; IL, intermediate lobe; PL, posterior lobe. Original magnification  $\times 100$  (A–C) and  $\times 400$  (D–I); bar = 50 µm (A–I).

Excel, respectively. All  $P$  values of less than 0.05 were considered to be statistically significant.

## RESULTS

### *Proliferation in the Pituitary of Females Is Increased Due to In Utero BPA Exposure*

Previous studies analyzing BPA exposure have shown increases in proliferation of a lactotrope pituitary cell line in vitro [28]. Embryonic exposure to BPA has also been shown to cause increased proliferation in endometrial glandular epithelial cells and prostate ductal epithelial cells in vivo [29,

30]. To identify if BPA exposure during embryonic development resulted in altered proliferation in the pituitary, pituitaries were collected from pups exposed to two low doses of BPA, 0.5 and 50 µg/kg/day, in utero. To examine proliferation, we used immunohistochemistry analysis and quantitative RT-PCR (qRT-PCR) to assess the expression of mKi67, which is present in cells that are actively progressing through the cell cycle. mKi67-immunoreactive cell number appeared to be greater in both the 0.5 µg/kg/day dose (Fig. 1, B and E) and the 50 µg/kg/day dose of BPA (Fig. 1, C and F) in females compared to the vehicle control (Fig. 1, A and D). In parallel, *mki67* mRNA levels showed significant increases in females of both treatment groups (Fig. 1J). Surprisingly,

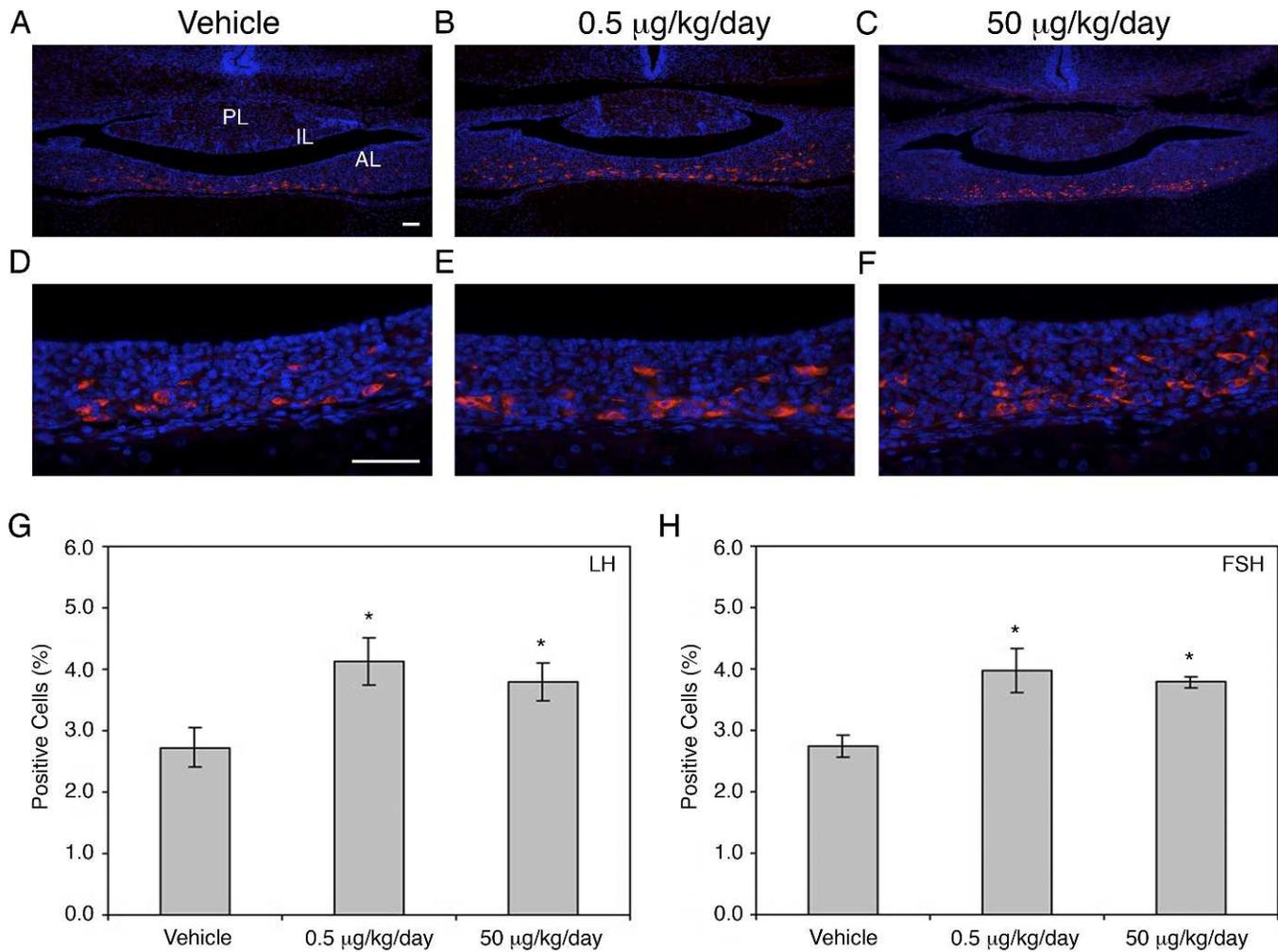


FIG. 2. LH $\beta$  and FSH $\beta$  positive cells were increased by in utero BPA exposure. Immunohistochemical detection of LH $\beta$  showed an increase in the pituitaries of both the 0.5  $\mu\text{g}/\text{kg}/\text{day}$  (B and E) and the 50  $\mu\text{g}/\text{kg}/\text{day}$  (C and F) BPA-treated mice compared to the vehicle control (A and D). LH $\beta$ -immunoreactive (G;  $P=0.029$ , 0.5  $\mu\text{g}/\text{kg}/\text{day}$ ;  $P=0.049$ , 50  $\mu\text{g}/\text{kg}/\text{day}$ ) and FSH $\beta$ -immunoreactive (H;  $P=0.020$ , 0.5  $\mu\text{g}/\text{kg}/\text{day}$ ;  $P=0.002$ , 50  $\mu\text{g}/\text{kg}/\text{day}$ ) cells are shown as the average percentage of positive cells compared to total cells in the anterior pituitary. Significance in G and H is noted by an asterisk ( $*P \leq 0.05$ ,  $n = 4$ ). AL, anterior lobe; IL, intermediate lobe; PL, posterior lobe. Original magnification  $\times 100$  (A–C) and  $\times 400$  (D–F); bar = 50  $\mu\text{m}$  (A–F).

mKi67 immunoreactivity did not appear to be altered in males between the vehicle control (Fig. 1G), the 0.5  $\mu\text{g}/\text{kg}/\text{day}$  dose (Fig. 1H), and the 50  $\mu\text{g}/\text{kg}/\text{day}$  dose of BPA (Fig. 1I), which is reflected in the comparable *mKi67* mRNA levels in all treatment groups (Fig. 1K).

#### Gonadotroph Cell Number Is Increased by In Utero BPA Exposure

Exposure to BPA during critical developmental periods has been shown to impact gonadotroph function. Therefore, we used immunohistochemistry for LH $\beta$  to visualize the effects of in utero exposure to BPA on gonadotroph cell number throughout the pituitary of females. LH $\beta$  immunostaining revealed an increase in gonadotroph number in both the 0.5  $\mu\text{g}/\text{kg}/\text{day}$  (Fig. 2, B and E) and the 50  $\mu\text{g}/\text{kg}/\text{day}$  BPA treatment groups (Fig. 2, C and F) compared to the vehicle control (Fig. 2, A and D). Quantification of the number of LH $\beta$ -immunopositive (Fig. 2G) and FSH $\beta$ -immunopositive (Fig. 2H) cells confirmed a significant increase in gonadotroph cell number.

#### Progenitor Proliferation Appears to Increase in Females Exposed to BPA

To identify if the increase noted in gonadotrophs was a direct result of an increase in their proliferation, immunohistochemistry was used to double-label mKi67 (nuclear, green) and FSH $\beta$  (cytoplasmic, red). Despite an increase of gonadotrophs in both the 0.5  $\mu\text{g}/\text{kg}/\text{day}$  (Fig. 3B) and the 50  $\mu\text{g}/\text{kg}/\text{day}$  (Fig. 3C) treatment groups compared to the vehicle control (Fig. 3A, solid arrow; for quantification, see Fig. 2), the majority of proliferating cells were not gonadotropin positive (arrowhead), although a few colabeled cells were detected in all groups (outlined arrow).

The pituitary-specific transcription factor PIT1 is necessary to produce somatotropes, lactotropes, and thyrotropes [21]. To identify if the proliferation of these PIT1-positive cell types was altered by BPA exposure, immunohistochemistry was used to double-label mKi67 (nuclear, green) and PIT1 (nuclear, red). Several double-labeled cells were detected (outlined arrow) in the control (Fig. 3D), 0.5  $\mu\text{g}/\text{kg}/\text{day}$  (Fig. 3E), and the 50  $\mu\text{g}/\text{kg}/\text{day}$  (Fig. 3F) treatment groups; however, the majority of the

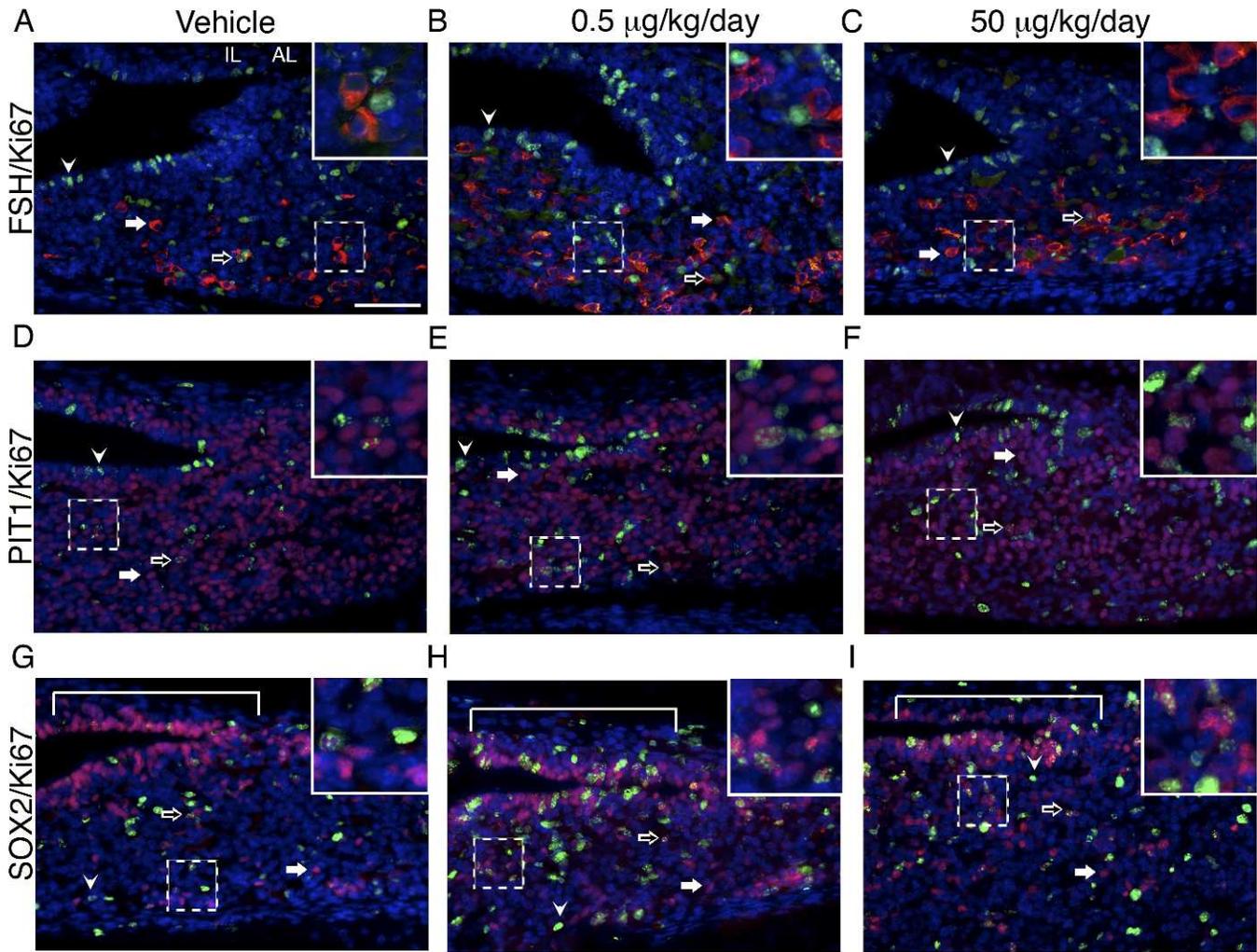


FIG. 3. BPA appeared to increase proliferation in pituitary progenitor cells. Few mKi67 (green) and FSH $\beta$  (red) colabeled cells were found in the female vehicle control pituitary (A), which does not appear to be changed by either the 0.5  $\mu\text{g}/\text{kg}/\text{day}$  dose (B) or the 50  $\mu\text{g}/\text{kg}/\text{day}$  dose (C) of BPA. Similarly, there appeared to be no difference in mKi67-immunoreactive (green) and PIT1-immunoreactive (red) cells, with a few double-labeled cells detected in all conditions (D–F). Many SOX2-positive progenitor cells in the cleft region (red) were double labeled with mKi67 (green) in both treatment groups (H and I, bracket), compared to only a few in the vehicle control (G, bracket). A representative cell immunoreactive with mKi67 only is denoted with an arrowhead in each panel. Cells immunoreactive only for FSH, PIT1, or SOX2 are shown with solid arrows, and outlined arrows indicate double-labeled cells. AL, anterior lobe; IL, intermediate lobe.  $n = 4-5$ . Original magnification  $\times 400$ ; bar = 50  $\mu\text{m}$  (A–I).

proliferating cells (arrowhead) were not PIT1 positive (solid arrow).

The well-characterized transcription factor SOX2 is essential for stem cell self-renewal and progenitor maintenance. In the pituitary, progenitor cells containing SOX2 are located in abundance near the cleft, lining the intermediate lobe of the pituitary and scattered throughout the anterior lobe [31, 32]. Because the localization of the proliferating cells in mice treated with BPA appeared similar to SOX2-containing cells, immunohistochemistry was used to determine if the amount of proliferating SOX2 cells increased. Cells were double labeled using mKi67 (nuclear, green) and SOX2 (nuclear, red). In the vehicle control group (Fig. 3G), although some SOX2-positive cells were colocalized with mKi67 (outlined arrow), many SOX2 cells were not proliferating (solid arrow). In both the 0.5  $\mu\text{g}/\text{kg}/\text{day}$  (Fig. 3H) and the 50  $\mu\text{g}/\text{kg}/\text{day}$  (Fig. 3I) treatment groups, the number of mKi67-positive cells appeared to be increased (arrowhead), and many of these cells colocalized with SOX2 (outlined arrow).

#### *In Utero BPA Exposure Alters Gonadotropin mRNA Levels and Factors Important for Gonadotroph Differentiation in Females*

Because an increase in gonadotroph cell number was detected in the pituitaries of BPA-treated female offspring at birth, we examined gonadotropin mRNA levels using qRT-PCR. Females exposed to 0.5  $\mu\text{g}/\text{kg}/\text{day}$  of BPA during embryonic development showed a significant increase in both *Lhb* (Fig. 4A) and *Fshb* mRNA (Fig. 4B), whereas males exposed to the same conditions showed no changes in expression (Fig. 4, A and B). However, females exposed to the higher BPA dose of 50  $\mu\text{g}/\text{kg}/\text{day}$  showed a significant decrease in mRNA levels of both *Lhb* and *Fshb*, whereas males exposed to the same conditions again showed no effects (Fig. 4, A and B). Because gonadotroph number was increased in both treatment groups but *Lhb* and *Fshb* mRNA levels were differentially altered, we examined mRNA levels of markers known to be important in gonadotropin synthesis. *Egr1* and *Nr5a1* (steroidogenic factor-1) are key components of gonadotropin synthesis [25, 26, 33]. GNRH receptor

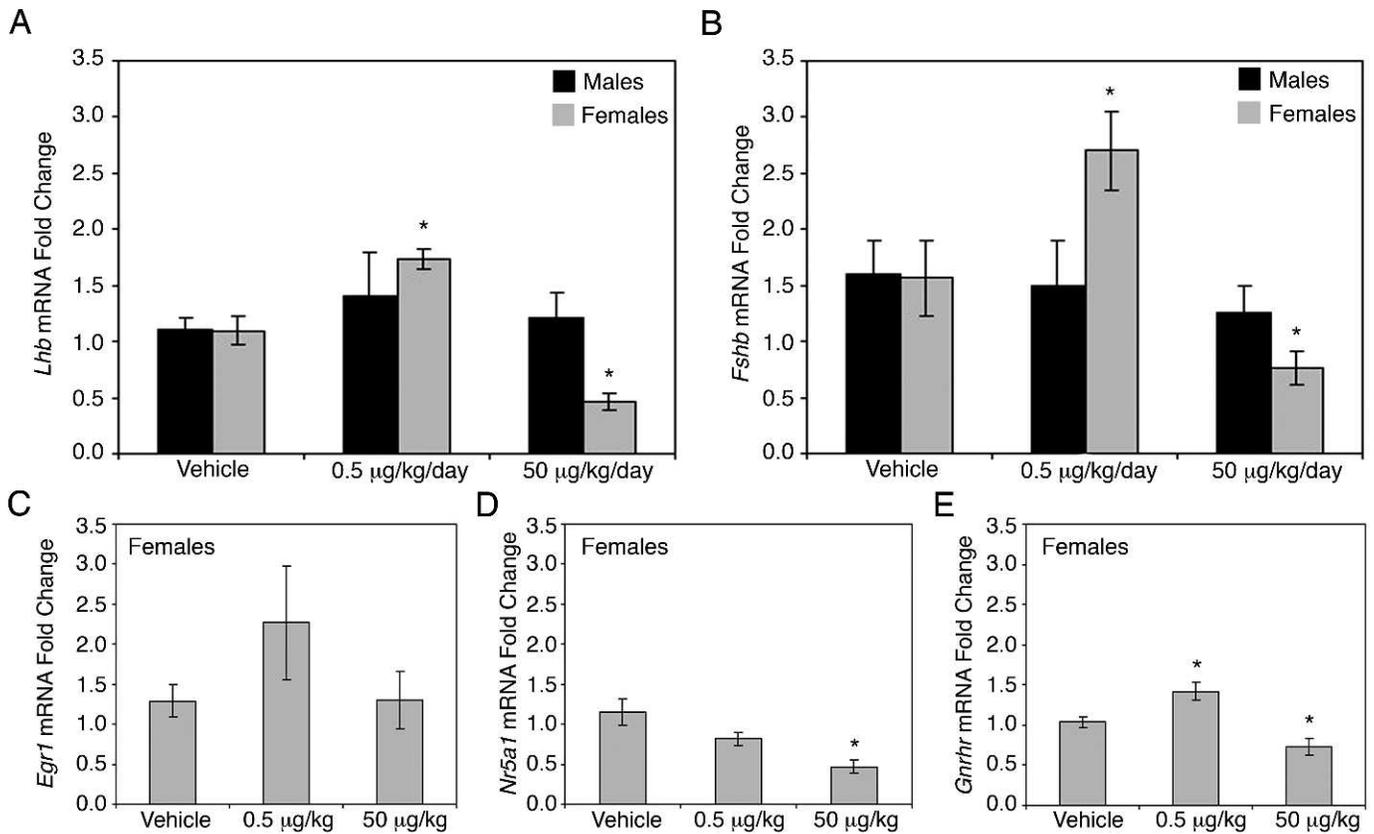


FIG. 4. Female gonadotropin mRNA levels and expression of factors controlling gonadotroph differentiation were altered by in utero BPA exposure. Levels of *Lhb* (A;  $P=0.004$ ) and *Fshb* (B;  $P=0.018$ ) mRNA were significantly increased in females exposed to BPA at a dose of 0.5 µg/kg/day compared to the vehicle control. Females exposed to 50 µg/kg/day of BPA showed a significant decrease in mRNA levels of both *Lhb* (A;  $P=0.002$ ) and *Fshb* (B;  $P=0.032$ ). No difference was found in the mRNA levels of *Lhb* (A) or *Fshb* (B) in males exposed to the same doses. In females, the mRNA levels for *Egr1* showed a trend towards an increase in the lowest treatment group ( $P=0.079$ ), but no change was observed for the 50 µg/kg/day treatment group (C). *Nr5a1* mRNA levels were significantly decreased in the 50 µg/kg/day treatment group (D;  $P=0.002$ ). *Gnrhr* mRNA levels showed a significant increase in the 0.5 µg/kg/day treatment group (E;  $P=0.011$ ) and a significant decrease in the 50 µg/kg/day treatment group (E;  $P=0.033$ ). All qRT-PCR values are normalized to the housekeeping gene *Gapdh* and are represented relative to the vehicle control group. Significance is noted by an asterisk ( $*P \leq 0.05$ ,  $n=7-8$ ).

(GNRHR) is essential for communication between the pituitary and the hypothalamus, as well as for initiating and maintaining gonadotroph hormone expression [34–36]. Although not significantly altered, *Egr1* showed a trend towards an increase in the lowest treatment group, but no changes were observed for the 50 µg/kg/day treatment group as compared to vehicle (Fig. 4C). *Nr5a1* had significantly decreased mRNA levels in the 50 µg/kg/day treatment group and no change in the 0.5 µg/kg/day treatment group (Fig. 4D). *Gnrhr* mRNA levels paralleled *Lhb* and *Fshb* mRNA levels, with a significant increase in the 0.5 µg/kg/day treatment group and a significant decrease in the 50 µg/kg/day treatment group as compared to vehicle treated controls (Fig. 4E).

#### Levels of PRL Protein and mRNA Are Undetectable at P1, and In Utero BPA Exposure Does Not Induce Expression

It is known that BPA can stimulate PRL production from lactotrope both in vivo and in vitro in the adult [28, 37–39] and can induce lactotrope proliferation [40, 41]. To determine if lactotropes were increased in response to BPA at P1, immunohistochemistry was used to examine PRL expression. No detection of PRL was found in any treatment group (Fig. 5, A–C). Additionally, in both male and female offspring at P1, *Prl* mRNA levels for all treatment groups were below the level of detection for qRT-PCR (data not shown). To determine

when in the course of development PRL was detectable, a time course using pituitaries of CD-1 mice collected at P1, P5, P10, and adulthood was immunostained for PRL. No PRL-positive cells were observed at P1 (Fig. 5, D and G), a small number of cells were visible at P5 (Fig. 5, E and H), many more cells were positive at P10 (data not shown), and a large amount of cells were positive by adulthood (Fig. 5, F and I). Because mRNA must turn on before protein is detectable, we used RT-PCR to characterize when *Prl* mRNA reaches detectable levels. Similar to the immunohistochemistry data, *Prl* mRNA levels were not detectable at P1, only present in low levels at P5, and easily detectable at P10 and adulthood (Fig. 5, J and K).

#### BPA Exposure Does Not Alter mRNA Levels of Major Hormones Produced by Somatotropes, Corticotropes, or Thyrotropes at P1

To elucidate if any other pituitary cell type was affected by in utero BPA exposure, the mRNA levels of hormones produced by corticotropes, somatotropes, and thyrotropes were analyzed. No changes were detected in the mRNA levels of pro-opiomelanocortin (*Pomc*), growth hormone (*Gh*), or thyroid-stimulating hormone (*Tshb*) for either male or female mice treated with BPA as compared to their vehicle-treated counterparts (Fig. 6).

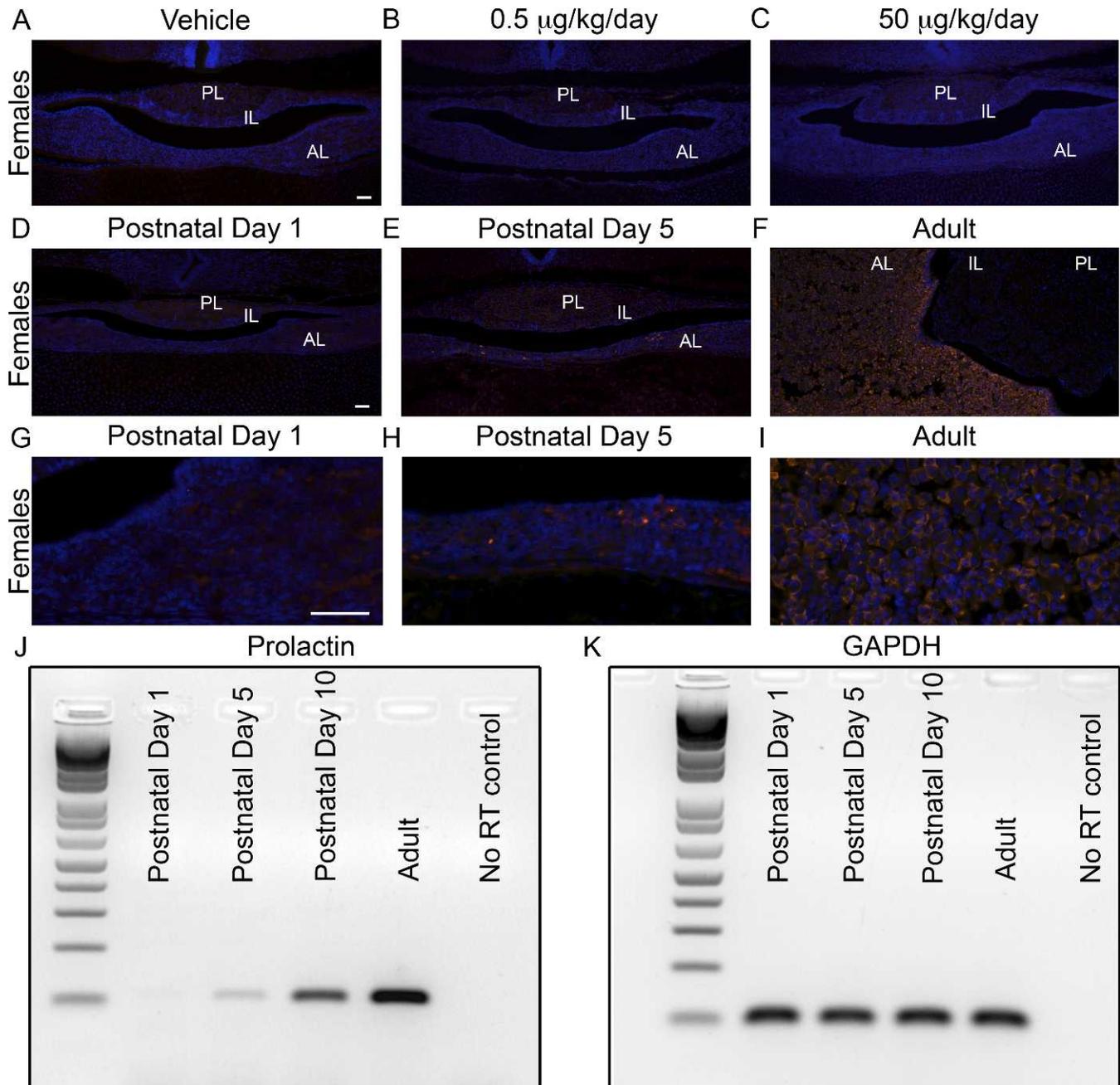


FIG. 5. PRL expression was not detectable at P1, and BPA exposure did not induce its expression. PRL-immunoreactive cells were not present at P1 in vehicle-treated female pituitaries (A) or those exposed to a 0.5 µg/kg/day (B) or 50 µg/kg/day (C) dose of BPA. A time course of PRL expression in CD-1 mice revealed no immunoreactive cells at P1 (D and G), a few cells at P5 (E and H), and many cells by adulthood (F and I). *Prl* mRNA was similarly not detectable in P1 pituitaries, and expression increased from P5 to adulthood (J). *Gapdh* is shown as a loading control (K).  $n = 4$  (Immunohistochemistry),  $n = 6$  (RT-PCR). AL, anterior lobe; IL, intermediate lobe; PL, posterior lobe. Original magnification  $\times 100$  (A–F) and  $\times 400$  (G–I); bar = 50 µm (A–I).

## DISCUSSION

Pituitary development relies on coordination between progenitor proliferation and differentiation of the distinct hormone-producing cell lineages (for review, see [4]). Failure to establish a pituitary with the proper cohort of cells at birth can result in hormonal imbalances and infertility. The extent to which the in utero environment can impact development of the pituitary is unknown. Using an in vivo model, we have shown that BPA exposure at environmentally relevant levels disrupts normal pituitary development. Female offspring are sensitive to

BPA exposure, and the gonadotrophs appear to be the main hormone-secreting cell type affected in the pituitary at birth. Furthermore, we have shown that the level of exposure is critical when examining the effects of BPA: Pituitary proliferation and gonadotroph number are affected similarly at both doses, but gonadotropin mRNA levels have a unique response in each dose.

Normally, pituitary progenitor cells will actively proliferate in response to signaling from the surrounding hypothalamic tissue during development [42, 43]. This helps to ensure appropriate number and specification of cells during organo-

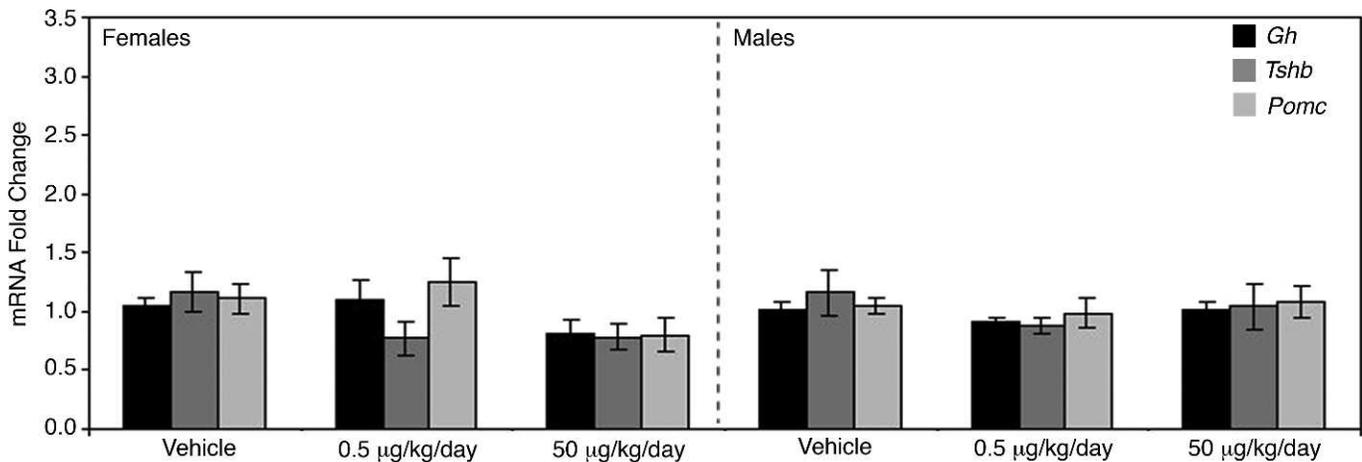


FIG. 6. The mRNA levels for the major hormones secreted by somatotropes, thyrotropes, and corticotropes showed no change in response to BPA exposure. The mRNA levels of hormones produced by somatotropes (*Gh*), thyrotropes (*Tshb*), and corticotropes (*Pomc*) of both males and females appeared unaffected by either dose of BPA tested. All qRT-PCR values are normalized to the housekeeping gene *Gapdh* and are represented relative to the vehicle control group.

genesis as well as to maintain stem cell populations in the adolescent and adult pituitary [44, 45]. The present results show that in utero exposure to BPA can increase pituitary proliferation. The mechanism by which BPA promotes pituitary proliferation is unknown; however, one possibility is that the action of BPA on the pituitary may be estrogenic. Developmental BPA exposure can increase proliferation in other estrogen-sensitive organs, such as the uterus and mammary gland [30, 46–48]. Furthermore, maternal exposure to estrogens or other estrogenic compounds, such as methoxychlor, has been shown to increase gonadotroph cell numbers [49, 50]. It is well established that the adult pituitary is highly sensitive to estrogen exposure, because estradiol treatment after a gonadectomy results in increased pituitary proliferation and exposure to estrogenic compounds, such as diethylstilbestrol, increases proliferation and promotes pituitary tumor formation [51, 52]. Besides estrogen receptors (ERs), BPA can act antagonistically with other hormone receptors, including thyroid hormone receptor (THR) and androgen receptor (AR) [53, 54], which are present in the embryonic pituitary [55, 56]. However, no evidence has been reported to date to indicate that activation or inhibition of THR or AR can influence pituitary cell number and specification during development. Additional studies are needed to address the mechanism of action of BPA at the level of the pituitary, but the results of the present study are consistent with a potential effect at ERs.

The present results show that exposure to BPA during embryonic development causes a significant increase in gonadotroph cell number in the pituitary. Interestingly, we did not observe any changes in mRNA levels of other pituitary hormones. These results are similar to those of other studies that have shown CD-1 mice exposed neonatally and through adolescence to BPA showed no changes in *Tshb*, *Gh*, or *Prl* mRNA [57]. However, it is somewhat surprising that no effect on lactotrope cell numbers was found, because in the adult pituitary, BPA exposure increases numbers of PRL cells in estrogen-sensitive rats (F344) in vivo and proliferation of somatolactotrope (GH3) cell lines in vitro [28]. In the present study, PRL protein and mRNA were too low to be detected in all of the treatment conditions, indicating that at the time point we examined, the initiation of PRL expression is not affected by BPA exposure. Overall, these data suggest that BPA can augment the proliferation normally occurring in the embryonic

pituitary and potentially direct uncommitted cells to the gonadotroph fate on the day of birth. This idea would be consistent with the studies showing that estradiol or methoxychlor can cause general proliferation increases and subsequent gonadotroph specification [49, 50]. Although the mechanism for this is unknown, it may be that intermediate cells are produced from the stem/progenitor cell that have many different receptors to sense the environment and differentiate depending on what signals are present.

Gonadotroph specification occurs through expression of developmentally regulated transcription factors and is reinforced by GNRH signaling late in gestation. Interestingly, despite seeing an increased number of gonadotrophs in response to both doses of BPA tested, the higher dose of 50 µg/kg/day was found to suppress *Lhb* and *Fshb* mRNA levels. We hypothesize that the increase in gonadotroph number may be established early in development and that the overall suppression of gonadotropin mRNA may be a result of a secondary action of BPA at the higher dose. The decrease in gonadotropin mRNA may be a result of a decrease in sensitivity of the gonadotrophs to GNRH, which is supported by the reduction seen in *Gnrhr* mRNA levels. GNRHR is expressed on the surface of pituitary gonadotroph cells and mediates the release of LH and FSH. It is also necessary for gonadotropin specification, because mice lacking *Gnrhr* have severely decreased LH $\beta$ -immunopositive cells and a decreased number of FSH $\beta$ -immunopositive cells [58]. In addition, we find *Nr5a1* (*Sfl*) mRNA levels are also decreased in the higher dose of BPA tested. *Nr5a1* is necessary for proper LH levels [26, 33]. These results indicate that although BPA can alter gonadotroph number, it may also impact *Lhb* and *Fshb* transcription through an unexplored mechanism. It is important to note that time of exposure and dose may be critical factors in determining the outcome of BPA administration. For example, adult exposure to BPA at higher levels than those used in the present study resulted in increases in *Lhb* and *Fshb* mRNA [57, 59], however, gonadotroph number was not examined.

One of the most significant findings of the present study is that embryonic BPA exposure only affects the female pituitary. One possible explanation is that sexually dimorphic development of the anteroventral periventricular nucleus (AVPV) of the hypothalamus is disrupted by BPA, leading to altered pituitary function. Males have higher circulating testosterone

concentrations than females in the late prenatal and early postnatal periods. Testosterone is aromatized to estradiol locally in the brain, which is critical in the masculinization of the AVPV [60, 61]. BPA administration after birth specifically affects female rats, leading to a reduction of *Kiss1* expression in the AVPV [62]. *KISS1* neurons are intimately involved in gonadotropin release from the pituitary at puberty, providing a potential link to the female-specific pituitary changes we see in response to BPA. However, the effect of exposure to BPA on the hypothalamus before birth has not been described, and it is unclear if the development of sexually dimorphic nuclei in the hypothalamus has any effect on prenatal pituitary development. Another possibility, suggested by human studies, is that the pituitary is sexually dimorphic as well, because gonadotroph numbers are greater in females than in males during early embryogenesis [63]. It is unknown, though, if the rodent pituitary is sexually dimorphic before birth or differentially exposed to estradiol during development. Evidence exists that postnatally, male pituitaries have higher levels of aromatase than female pituitaries [64], similar to the hypothalamus, but again, if this is also true during embryogenesis is unknown. Although BPA exposure has been shown to have a sex-specific effect in the brain, the nature of its mode of action in developing female pituitaries will clearly warrant further investigation.

Understanding the consequences of BPA exposure on the developing pituitary is critical because of its widespread use and evidence for long-term reproductive effects of prenatal exposure (for review, see [65–68]). The present study defines the effect of BPA exposure during the critical period of pituitary organogenesis. We demonstrate the potential for low, environmentally relevant doses of BPA to have an effect on gonadotroph cell number and specification in female pituitaries. This knowledge is a critical step in furthering our understanding of how the pituitary responds to exogenous chemical exposure during development and suggests that the ORfD set by the government may be too high for sensitive subgroups, such as pregnant women.

## ACKNOWLEDGMENT

The authors thank Dr. Ann Nardulli at the University of Illinois Urbana-Champaign for use of equipment and all members of the Raetzman Lab for technical support and reading of the manuscript.

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